



Multivariate statistical evaluation of dissolved trace elements and water quality assessment in the Karaca dam, Turkey

Banu Kutlu¹, Ekrem Mutlu²

Munzur University, Faculty of Fisheries, Tunceli, Turkey
 Kastamonu University, Faculty of Fisheries, Kastamonu, Turkey

Corresponding author e-mail: <u>banukutlu@munzur.edu.tr</u>

ARTICLE INFO

Received 15/1/2021; received in revised form 5/4/2021; accepted 30/4/2021. DOI: <u>10.6092/issn.2281-4485/12231</u> © 2021 The Authors.

Abstract

This study was performed from January 2017 to December 2017 with taking samples every month from four stations to determine the water quality of Ka-raca Dam, located in Sinop. The guidelines of the World Health Organization for water quality index (WQI) and Turkey's Ministry of Forestry and Water Affairs specifications for Surface Water Quality Regulations (SWQR) were used in the determination of water quality and 28 variables were analyzed. The quality of the irrigation water was also calculated. This objective, the sodium adsorption ratio (SAR), sodium percentage, permeability index (PI), and mag-nesium index were calculated. The average WQI value of the lake was found to be 19.23. It was observed that the water quality parameters did not exceed the restrictions in any stations during the period. According to the results of the research, Karaca Dam is in a good class in terms of drinking water quality and irrigation water. The health risk evaluation, using the hazard ratio and index (HI), was carried out as suggested by USEPA, and it was found that the water of Karaca Dam is not harmful to human health in terms of Pb, Cd, and Hg.

Keywords

Karaca dam, water quality, water quality index, irrigation water quality, surface water

Introduction

Of the world's water body, 97.6% is composed of oceans, and 2.4% is fresh water. Only 1% of freshwater is suitable for human consumption. Dams and artificial dams are built to improve freshwater content. Recently, there has been concern about the shortage of water resources in the future (Ongley, 1998). The studies are aimed at meeting the water quality expectations of streams and rivers, as well as protecting the groundwater and surface water resources (Banerjee and Srivastava, 2009). The persistence of freshwater resources is important for the quality of life on earth (Manjare et al.,

2010). However, water resources get affected by several contaminants in today's world. These cause changes in the physicochemical quality of water (Subramani et al., 2009). These contaminants are of industrial, agricultural, and domestic origin (Kumar and Thakur, 2017a,b; Kumar et al., 2015; Yu et al., 2015; Wu et al., 2018). Agricultural lands are easily contaminated by wastewater and soluble contaminants. Industrial wastewater, agricultural land nutrients, organic, and inorganic wastes are the main contaminants whose precise source is unknown (Ali and Khairy, 2016).

Also, nitrogen and phosphorus cause contamination by leaching to the surface water (Smith 2016). For this reason, monitoring programs are developed, which are aimed to maintain the water quality and determine water quality parameters (Mullai et al., 2013; Phung et al., 2015; Yisa and Jimoh, 2010).

High concentrations of heavy metals in water and earth directly affect human health (Mortuza and Al Misned 2017). Heavy metals are another contaminant of the aquatic system and human health (Rupakheti et al, 2017). It is known that anthropogenic activities accelerate the accumulation of heavy metals in water, which eventually deposit in the human body through the food chain. Anions and cations, such as NO_3 , PO_4 , SO_4 , Ca, Mg, Fe, and inorganic salts, are the main contaminants in water (Zafar et al., 2017; Anwar et al., 2017; İmran et al., 2017). In drinking and irrigation waters, Na, Ca, Mg, and K are necessary for organisms to live. Although trace amounts of copper and zinc are important for life, their high concentrations cause toxic effects (Bing et al., 2013).

One of the most common methods to determine the quality of water is the water quality index (WOI), which was used by Horton (1965) (Tripathi and Singal, 2019). It is crucial for the balance of the ecosystem and public health to monitor the nutrients, metals, and physical qualities of water resources. Dams are important resources, which must be protected to increase the availability of water for human consumption, and maintain the sustainability for future generations. The quality of dam water, which is located around

minerals, agriculture, and animal husbandry in the Sinop province, was evaluated using the water quality index and multivariate statistical approaches in this study, the physicochemical parameters of surface waters and water quality parameters, including nutrient and heavy metal concentrations of important irrigation dams in the Sinop region, were investigated. Using the obtained data, the surface water contamination value of Karaca Dam was determined by the WOI index. The relationship between contamination sources and their varieties were identified using Turkey's Ministry of Forestry and Water Affairs surface water quality regulations, and multivariate statistics. Accordingly, the potential usage of the water for drinking and irrigation purposes was investigated.

Materials and Methods

Karaca Dam Lake is located in part of the Kızılırmak Basin in the province of Sinop. The water source of the dam is Karacaören brook. The dam has an area of 0.144 km², and the capacity of the lake is 1.817 hm³. Although the dam is located in the hunting area of Sinop province, it is within the wildlife protection area and meets the water needs of wild animals. The study was carried out every month in 4 stations for 12 months. From January 2017 to December 2017, surface water sampling was carried out monthly at four stations considered to represent the entire lake (Fig. 1). The water samples were taken in acid-cleaned 2.5L sampling bottles from 15 to 20 cm below the water surface.



Figure 1. Location of sampling stations in Karaca Dam Lake.

EQA 44 (2021): 26-31

The water samples for heavy metal analysis were taken in polyethylene bottles, which were previously washed with 50% HNO₃ and deionized water and acidified with 10 mL HNO₃ per liter. The samples were transported to the laboratory in iceboxes, and stored in the refrigerator at 4 °C until analysis. Water temperature, dissolved oxygen (DO), pH, salinity, and electrical conductivity (EC) values were measured in-situ with YSI 556 MPS. Chemical oxygen demand (COD), biological oxygen demand (BOD), total hardness (TH), nitritenitrogen (NO₂²⁻-N), nitrate-nitrogen (NO₃²⁻-N), ammonium-nitrogen (NH₄⁺-N), total alkalinity (TA), orthophosphate (PO_4^{3-}), sulfite (SO_3^{2-}), sulfate (SO_4^{2-}), chloride (Cl⁻), calcium (Ca²⁺), magnesium (Mg²⁺), sodium (Na⁺), and potassium (K⁺) analyses were performed using standard methods (APHA, 1998). Iron, lead, cadmium, zinc, nickel, copper, and mercury analyses were performed using the PerkinElmer Optima 2000 DV ICP-OES.

Water quality index (WQI)

Water quality index (WQI) is defined as a grading technique that measures the combined effect of each of the water quality parameters to evaluate the overall water quality for human consumption (Horton 1965; Kangabam et al., 2017). It has been used extensively in recent years to analyze the potential of water resources to be used for drinking and domestic purposes (Khanoranga, 2019). WQI based on the nineteen parameters (pH, EC, BOD, Cl, SO₄, Na, K, Total Hardness, Ma, Ca, NO₂, NO₃, Fe, Pb, Cu, Cd, Hg, Ni, and Zn) that are important for the analysis of water quality was determined using their seasonal and average values. An actual weight (AW) between 1 and 5 was assigned to each parameter, depending on its effect on the water quality, and its significance on human health. The relative weight (RW) value was calculated by the following formula (Eq. [1]).

$$RW = \frac{AW}{\sum_{i=1}^{n} AW}$$
[1]

After calculating RW, each of the analyzed parameters is divided by drinking water values (S_i) permitted by the World Health Organization (WHO, 2011), and then multiplied by 100 to calculate the quality rating (Qi) (Eq. [2]). Whereas "Ci" is the observed concentration of each parameter and "Si" are WHO standards.

$$Q_i = \left(\frac{C_i}{S_i}\right) \times 100$$
[2]

Later, the sub-indices (SI) were calculated (Eq. [3]) and added to determine WQI (Eq. [4]).

$$SI_i = RW \times Q_i$$
 [3]

////

$$WQI = \sum_{i=1}^{n} SI_i$$
[4]

According to WQI, it is further evaluated by

$$SAR = \frac{Na_{meq}^{+}}{\sqrt{\frac{Ca_{meq}^{2+} + Mg_{meq}^{2+}}{2}}}$$
[5]

Sodium percentage (%). Evaluation of sodium percentage (%) is important because high sodium rates in irrigation waters cause stunted growth in plants. For this reason, the sodium rate was calculated (Ghalip, 2017).

$$\% Na = \frac{\left(Na_{meq}^{+} + K_{meq}^{+}\right) \times 100}{Na_{meq}^{+} + Ca_{meq}^{2+} + Mg_{meq}^{2+} + K_{meq}^{+}}$$
[6]

Magnesium hazard (MH). Generally, Ca^{2+} and Mg^{2+} are present in equal amounts in water. Mg^{2+} is present in high amounts in some water sources and controls its alkalinity, which affects the growth of plants. The Mg concentration (MH) was determined using the following equation (Abdulhusseyin, 2018; Khanoranga, 2019).

$$H = \left(\frac{Mg_{meq}^{2+}}{Ca_{meq}^{2+} + Mg_{meq}^{2+}}\right) \times 100$$
 [7]

Permeability index (PI). If water, which is rich in minerals, is continuously used for irrigation, the transmittance of the earth decreases. Permeability index was calculated as follows (Falowo et al., 2017).

 $PI = (Na^+ + \sqrt{HCO3}) * 100/(Ca2 + +Mg2 + +Na+)$ [8] The factor, correlation, and clustering analyses were performed to determine the processes occurring in the ecosystem, identify the relationship between calculated values, and relate the sources.

Results and Discussion

The physicochemical parameters of the irrigation dam showed the natural quality of the water (Khanoranga, 2019). General qualities of the physicochemical parameters are given in Table 1.

-	e e						
Parameter	Winter (n = 12)	Spring (n = 12)	Summer (n = 12)	Autumn (n = 12)	Mean ±SD (n = 48)	Min	Max
DO (mg/L)	12.13ª	12.87 ^b	11.06°	11.04°	11.77 ±0.89	10.22	13.20
Salinity (‰)	0.10ª	0.27 ^b	0.42°	0.28 ^b	0.27 ±0.13	0.10	0.50
pН	8.31ª	8.46 ^b	8.66 ^c	8.50 ^b	8.48 ±0.16	8.25	8.78
T (°C)	5.15ª	10.34 ^b	19.58°	17.28 ^c	13.09 ±6.54	3.30	23.70
EC (µS/s)	237.11ª	241.88ª	291.77 ^b	299.34 ^b	267.52	203.50	317.66
TSS (mg/L)	0.60ª	0.92ª	2.19 ^b	1.81 ^b	1.38 ±0.76	0.35	2.76
COD (mg/L)	0.99ª	1.20ª	2.14 ^b	2.06 ^b	1.60 ±0.57	0.83	2.54
BOD (mg/L)	0.46ª	0.31ª	0.98 ^b	1.04 ^b	0.70 ± 0.38	0.12	1.34
CI (mg/L)	2.84ª	2.79ª	3.01ª	3.79 ^b	3.11 ±0.59	2.15	4.70
$PO_4 (mg/L)$	0.21	0.12	0.09	0.18	0.15 ±0.18	0.00	0.98
$SO_4 (mg/L)$	16.57ª	29.16 ^b	53.70°	41.99 ^d	35.36 ±14.95	14.77	59.26
$SO_3 (mg/L)$	0.36ª	0.69 ^b	1.17 ^c	0.81 ^b	0.76 ±0.32	0.29	1.30
Na (mg/L)	34.03ª	49.12 ^b	42.65°	29 ^d	38.70 ±8.85	26.52	53.42
K (mg/L)	5.40 ^{ab}	5.10ª	6.13 ^b	4.61ª	5.31 ±0.99	4.02	7.10
TH (CaCO ₃ mg/L)	185.77ª	239.47 ^b	218.44°	180.82ª	206.13	161.64	244.02
TA (CaCO ₃ mg/L)	187.40^{a}	241.13 ^b	219.18 ^c	182.72ª	207.61	161.94	245.66
Mg (mg/L)	23.71ª	30.94 ^b	36.90°	26.71ª	29.56 ±5.99	19.52	39.68
Ca (mg/L)	27.16ª	50.03 ^b	57.69°	44.23 ^d	44.78 ±12.38	23.40	59.02
$NO_2 (mg/L)$	0.00003ª	0.00007^{b}	0.00006^{b}	0.00005ª	0.00005	0.00001	0.00013
$NO_3 (mg/L)$	0.47a	0.70 ^b	0.74b	0.66 ^b	0.64 ±.13	0.38	0.85
NH_4 (mg/L)	0.00003ª	0.00003ª	0.00011^{b}	0.00013^{b}	0.00007	0.00001	0.00024
Fe (µg/L)	1ª	1.58 ^{ab}	2^{bc}	2.93°	1.88 ± 1.08	1.00	6.00
Pb (µg/L)	0.44^{a}	0.96 ^b	0.43ª	0.19ª	0.50 ±0.38	0.10	1.50
Cu (µg/L)	1.59ª	8.83 ^b	4.25°	4.50°	4.79 ±3.30	1.00	14.00
Cd (µg/L)	0.23	0.20	0.23	0.25	$0.23 \pm .08$	0.10	0.40
Hg (µg/L)	0.001ª	0.007^{b}	0.006 ^b	0.006	$0.005 \pm .003$	0.001	0.013
Ni (µg/L)	2.75ª	4.17 ^b	4.58 ^b	6.50°	4.50 ±1.61	2.00	8.00
Zn (µg/L)	3.83ª	16.17 ^b	10.75°	10.92°	10.42 ±6.41	2.00	25.00

 Table 1. Descriptive statistics of the variables
 Particular

The mean difference is significant at the 0.05 level.

The pH value of Karaca Dam is within the restrictive limits specified by WHO (2011), and SWQR, which are 6.8, and 8.5, respectively. According to SWOR (2016), the dam water is of good quality (Rasool et al., 2017; Ali et al., 2017, Arshad and Imran, 2017), and has sustainable pH (Tables 2 and 3).

Parameters	WHO 2011	Assigned weight (AW)	Weight Relative (RW)
pН	6.5-8.5 (7.5)	4	0.056
EC (µS/cm)	1500	4	0.056
BOD (mg/L)	5	5	0.070
Cl (mg/L)	250	4	0.056
$SO_4 (mg/L)$	250	5	0.070
Na (mg/L)	200	3	0.042
K (mg/L)	12	2	0.028
TH (mg/L CaCO ₃)	100	1	0.014
Mg (mg/L)	50	2	0.028
Ca (mg/L)	75	2	0.028
$NO_2 (mg/L)$	3	5	0.070
$NO_3 (mg/L)$	50	5	0.070
Fe (µg/L)	300	4	0.056
Pb (µg/L)	10	5	0.070
Cu (µg/L)	2000	2	0.028
Cd (µg/L)	3	5	0.070
Hg (µg/L)	6	5	0.070
Ni (µg/L)	70	5	0.070
Zn (µg/L)	3000	3	0.042

Table 2. Assignment of relative weight to the studied groundwater quality parameters.

According to WHO (2011), the EC values of Karaca Dam are within acceptable limits. The presence of contaminants might cause an increase in the EC value of surface waters (Şener et al., 2017). An increase in the EC hampers ionic absorption by the plants resulting in physiological drought (Naseem et al., 2010). A high value of EC could be due to animal farms, agricultural wastes, sewage, and discharge wastes (Kanhabam et al., 2017). Additionally, our study found that the EC values increase with the rise in temperature. Upon evaporation, the salinity increases, resulting in a rise in EC (Jiang et al., 2015; Zhang et al., 2016). According to the SWQR classification, the dam is first class in terms of EC value and included in the clean water class

The chemical oxygen demand (COD) is one of the important parameters used to measure the contamination of domestic and industrial wastewater. The fact that the quality increased during the autumn season suggests that it may be caused by the chemical contaminants in agricultural lands (Imneisi and Aydın, 2016). The biological oxygen demand (BOD), on the other hand, increased in December, whereas it decreased in March. This is due to the anthropogenic activities along with domestic waste and fish activities (Sallam and Elsayed 2018). The COD and BOD values are inversely related to DO (Table 4). In Karaca Dam, the concentration of sodium (Na) was within limits prescribed by the WHO (2011). The cationic exchange is associated with the geological characteristics of the solution of lithogenic sodium (Guo et al., 2007; Rafigue et al., 2008). The results show parallelism with our study. High cation rates might be the result of the change in the concentration of calcium and sodium on earth. The calcium concentration is generally low in natural water but is sufficiently high in drinking water, industrial, and irrigation waters (Kutlu et al 2017; Tepe and Kutlu, 2019). Calcium concentrations are within limits suggested by the WHO (2011). Similar results were obtained by Rapan et al., (2017) and Alam et al., (2017). The magnesium concentration is within the WHO (2011) terms. It is most commonly found in mineral rocks and seeps into the water through natural or anthropogenic ways. Its high rate negatively affects human health (Daud et al., 2017; Rasol et al., 2017). Potassium concentration is within limits specified by the WHO (2011). It is supposed to have come from potassium-containing rocks and agricultural fertilizers (Mumtaz et al., 2017). In the study of the Karaca Dam,

sulfate was found to be the highest in September and lowest in March, although it did not exceed the WHO (2011) values. The primary source of sulfate is the degradation of sulfide compounds by bacteria and the use of sulfated fertilizers (Yamamura, 2008, Varol and Davraz, 2015).

According to SWOR (2016), it was observed that the phosphate values did not exceed the limits (Class II). The agricultural fertilizers might be the cause of the presence of phosphate. The domestic and organic wastes increase the phosphate levels and cause eutrophication, which results in excessive levels of toxic algae (Boyd and Tucker, 1998). The nitrite concentration was lowest during other seasons compared to winter and autumn, and reached the maximum in spring. Nitrate concentration also shows increase and decrease in the same manner. Nitrite is a by-product of the reaction between ammonium and nitrate. Compared to others, it has a toxic effect on animals (Belal et al., 2016). The concentration of nitrate present in the water in our area is much lower than those in other places (Soomro et al., 2017). It is within the limits of the WHO (2016) and does not pose a threat to human health (Debels et al., 2005). Ammonium ions are produced during the biological degradation of organic matter and directly used by plants. Ammonium sulfate is commonly used by farmers as both organic and inorganic fertilizers (Vega et al., 1998). Fertilizers leach into the lakes and dams via the overflow of irrigation water.

In this study, the heavy metal concentrations, apart from Fe, were found to be over the limit suggested by the WHO (2011) (Table 2). It is thought that the reason for the high rates of these metals might be due to agricultural activities, mineral deposits, and anthropogenic wastes along with rocks (Kumari et al., 2017). The results show parallelism to other studies (Fatmi et al., 2009; Podgorski et al., 2016). Heavy metals accumulate in the liver, kidneys, and muscle tissues of aquatic organisms. Secondary sources of heavy metals are atmosphere and underground water (Ali et al., 2017). It is thought that the high concentration of heavy metals identified in the dam is a result of natural or anthropogenic activities.

Sodium absorption ratio (SAR). The values of sodium absorption ratio (SAR) are used to determine the water quality of irrigation water. The SAR value of Karaca Dam is 1.11 mEq/L (Table 4). If the SAR value is >9, the sodium level is hazardous and water cannot be used for irrigation (Awais et al., 2017). High SAR values might cause a decrease in salinity values and result in a decrease in the calcium and magnesium levels, which are essential elements for the growth of plants (Vasantvigar et al., 2010; Rasool et al., 2016). The result of the study shows that Karaca Dam water is suitable for irrigation.

Table 4. Water quality classification based on Na%, SAR, MH and WOI.								
Na %	Na % Water quality		SAR (meq/L) Water quality		Water quality	WOI Range		
< 20	Excellent	0–6	Good	< 50	Suitable	< 50	Excellent	
20-40	Good	6–9	Doubtful	> 50	Unsuitable	50-100	Good	
40–60	Permissible	> 9	Unsuitable			100-200	Poor	
60-80	Doubtful					200-300	Very Poor	
> 80	Unsuitable					> 300	Water unsuitable for drinking purpose	

Sodium percentage (%). In the study, the average percentage of sodium was found to be 36.18% (Table 4). Water has been classified based on the sodium percentage by several researchers (Srinivas et al., 2017; El Aziz, 2017; Islam et al., 2017). According to Table 4, the dam water belongs to the good quality category of water. Its high rate in water might be due to the use of fertilizers in agriculture. The high level of sodium (9%) in the land negatively affects the tightness, ventilation, and texture of the soil (Singara et al., 2014). The plants

treated with water with high sodium showed a low fertility rate because sodium affects the osmotic pressure between the soil and plant. Thus, it blocks the mineral uptake by the plant from soil (Nasseem et al., 2010).

Permeability index (PI). The permeability of the soil is affected by the presence of excessive minerals (Ca, Mg, Na, HCO_3) in water, and greatly affects long-term agricultural productivity (Sing et al., 2008). The permeability index negatively affects the fertility of soil

(Obiefuna and Sheriff, 2011). The use of water that is rich in minerals prevents the growth of seedlings by blocking the ventilation in soil. This index was developed by Doneen (1975) to classify water. According to this index, the water that has PI > 75 is the first class water, that with PI = 25–75 is the second class water, and that with PI < 25 is the third class water (Doneen, 1975; Raju et al., 2011). In our study, the permeability index (PI) was found to be 59.95; thus the region is in the second class water category (Table 4).

Magnesium hazard (MH). In our study, the average magnesium was found to be 40.39 mEq/L (Table 4). The classification of water in terms of magnesium hazard is related to its suitability for agricultural purposes. The high concentration of magnesium in water is a factor that threatens the fertility of soil (Kharonga, 2019). This is because high magnesium and clay pieces prevent the absorption of water (Hussain et al., 2016a, 2016b). The value detected in our study was found to be lower compared to those in other countries (Patet et al., 2017; Golekar et al., 2017; Padhi et al., 2017).

Water quality index (WOI). The water quality index (WOI) was employed to determine the water quality of the Karaca Dam. To calculate the index, Cd, Cl, Cu, Pb, Hg, Ni, nitrate, Na, bitterness, pH, SO₄²⁻, and Zn concentrations were determined. The WHO (2011) limit, as given in Table 2, was used as reference. The WOI values were analyzed for each season (rainy season–dry season) and were at the lowest level during winter (16.47) and at the highest level in summer (21.70). The suggested limit of water quality parameter did not exceed in any of the stations. According to these values, the Karaca Dam was found to have the perfect water quality.

Table 5. Water classification grading based on the

 Water Quality Index (WQI).

WOI	Season Mean
16.47	Winter
19.55	Spring
21.70	Summer
19.23	Autumn

According to the Cluster Analysis, it was detected that Chlorine (Cl), Potassium (K), Copper (Cu), and Nickel (Ni) have a close relationship with each other and mostly originate from common sources. The salinity, concentrations of Cd, PO_4 , NO_2 , NH_4 , Fe, SO_3 , NO_3 , Pb, TSS, BOD, and COD are closely related. This might be the result of their origin from the same source. The DO, temperature, pH, and zinc analyses formed a group together. SO_4^{2-} , Ca, Na, and Mg formed another group. The cationic group in the water created equalization. TA, TH, and electrical conductivity constituted a separate group (Fig. 2).



Figure 2. Dendrogram obtained by the cluster analysis.

The eigenvalues of the four factors were > 1. The total variables of these four factors were calculated to be 88% (Table 6). The first factor was 42%. While primarily TSS, COD, WT, SO₄, EC, BOD, SO₃, NH₄, salinity, pH, Fe, Ca, and Ni were positive, DO was negative. The rate was calculated to be 42%. Ammonium and temperature are in the first factor; this shows that they are the two important indicators of primary concern (Kutlu et al., 2017). DO is in negative correlation with the first factor (Table 6).

DOI: 10.6092/issn.2281-4485/12231

Parameter	Factor 1	Factor 2	Factor 3	Factor 4
TSS	.985	.035	.057	030
COD	.976	.126	073	001
WT	.954	.222	.090	045
SO_4	.938	.225	.047	.217
EC	.926	.021	248	115
BOD	.911	109	337	080
SO ₃	.907	.254	.238	.194
NH_4	.899	115	058	357
Salinity	.852	.285	.339	.124
pН	.850	.133	.455	.024
DO	837	.434	.236	.107
Fe	.714	.253	.027	477
Ca	.675	.460	.490	.184
Ni	.634	.502	333	352
Zn	.115	.946	.226	007
Cu	017	.940	.298	079
NO ₂	.071	.925	.070	.205
Hg	.316	.910	.165	.051
NO ₃	.485	.808	.088	.179
Pb	384	.632	.561	.075
TA	.030	.374	.884	.064
TH	.034	.376	.884	.088
Cl	.283	.122	816	.042
Na	180	.464	.765	.352
Mg	.624	.192	.671	.274
PO_4	157	037	566	.015
ĸ	.120	.119	.347	.633
Cd	.421	344	.299	532
Eigen values	12.703	7.874	3.046	1.273
Variability (%)	42,034	22,213	18.838	5.831
Cumulative%	42,034	64,247	83.084	88.915

Table 6.	Factor	loading	matrix	after	varimax	rotation.
----------	--------	---------	--------	-------	---------	-----------

As EC increased, especially in summer, the level of oxygen decreased. It was observed that in the primary production, among the nitrate forms, ammonium was preferred by phytoplankton. According to, Tepe and Kutlu (2019), while temperature, pH, and EC increase during summer, DO decreases. Such levels of BOD and COD in the lake water are due to excessive input of organic materials. Organic materials in lakes result from domestic wastes and the use of chemicals and fertilizers. Ammonium sulfate fertilizers are commonly used in

agriculture (Vega et al., 1998), which have probably leached into the lake along with agricultural discharge and organic wastes. The second factor was 2.22% of the total change and contained NO_3 , NO_2 , Zn, Cu, and Pb.

Factor 3 (F3) showed a cumulative positive effect at the rate of 18.83%, which includes the total bitterness, TH, and concentrations of Cl, Na, Mg, and PO_4 (Table 6). These are the result of the discharge of agricultural wastes into the Karaca Dam, and excessive use of

fertilizers (Kıymaz, 2010; Kıymaz and Karardavut, 2014). In the third factor, the presence of phosphate shows the effect of fertilizers and organic wastes. It was understood from the non-opposite directions that the sources of K and Cd are unnatural rocks. Pesticides contain nitrogen and phosphorus, which contribute to agricultural fertilization. Sugar beets, grains, barley, and corn are generally produced in the lake basins. Besides, the flow of phosphorus from agricultural lands to lake basins causes contamination of the lake.

Pearson's correlation test showed a negative correlation for pH, WT, EC, TSS, and COD. Also, the values of pH, WT, EC, SSD, BOD, NO, Mg, Ca, NH_4 , Fe, and Ni showed a positive correlation with each other. There was a positive correlation between NO_3 and NH_4 , as well. Na, Ca, and HCO_3 are the dominant ions in natural water. Similarly, there was a positive correlation between heavy metals and trace elements. Major sources of these heavy metals were the phosphate and inorganic nitrate fertilizers used in agricultural lands (Table7).

Table 7. Results of correlation analysis of the physiochemical parameters of the dam lake of study

	DO	Salinity	pH	WT	EC	TSS	COD	BOD	a	P04	SO.	50,	Na	ĸ	тн	ТА	Mg	Ga	NO,	NO,	NH,	Fe	Pb	Cu	Cd	Hg	N	Zn
DO	1																											-
Salinity	.48"	1																										
pH	.52"	.93"	1																									
WT	.69"	.90"	.88"	1																								
EC	.83"	.69**	.66**	.88"	1																							
TSS	.79"	.87**	.89**	.96**	.89**	1																						
COD	.77"	.83"	.81**	.95"	.90"	.96**	1																					
BOD	.89"	.61**	.61"	.80"	.90"	.88"	.90"	1																				
a	.32'	.03	.10	.16	.47"	,20	.33"	.49"	1																			
PO,		.27	.35*	.18	.02	.14	.17	.06	.34"	1																		
SO,	.64"	.91**	.83**	.93**	.85"	.92**	.93**	.78**	.29*	.16	1																	
SO	-36"	.97"	.91"	.93"	.76**	.92**	.89**	.70"	.11	.24	,97"	1																
Na	.58"	.31*	.29"	.01	.36"	.12	.18	.52 ^{**}	-54°	.37"	.05	.21	1															
к	.10	.23	.29*	.10	.04	.07	,12	.05	.18	.28	.25	.28*	.50"	1														
тн	.35'	.47**	.48**	.19	.19	.10	.01	-31*	.63**	.42**	.19	.37"	.88"	.27	1													
TA	.35'	.47**	.47**	.18	-20	.10	.00	.31*	.63**	.43**	.18	.36"	.87**	.25	.99"	1												
Mg	-26	.82"	.82**	.68**	.39"	.63**	.58"	.29"	-36"	.45"	.73**	.82**	.55"	.50"	.72"	.71**	1											
Ca	323	.90**	.84**	.78**	.52"	.69**	.66**	.35"	.09	-36"	.82**	.88"	.55"	.37"	.67**	.66**	.90"	1										
NO ₂	.36"	.36*	.20	.25	.01	.10	.18	04	.04	.10	.31*	.36*	.50**	.26	.43**	.43**	.33*	.50**	1									
NOs	.03	.69**	.53**	.64"	.40**	.50**	.57**	.32*	.10	.18	.68**	.71**	.38"	.23	.42**	.41**	.57**	.75**	.87**	1								
NHe	.85"	.66**	.72**	.82**	.84**	.89**	.88"	.90"	.26	.13	.72**	.69**	.40"	.13	.09	.08	.39"	.42**	08	.28	1							
Fe	-s1"	.60**	.70**	.76**	.67**	.73**	.74**	.66**	.17	14	.58"	.59"	.14	.02	.05	.06	.34"	.48"	.16	.41**	.82**	1						
Pb	.76**	.04	.03	.20	.46**	.33*	-34"	.61**	.43**	.27	.17	.04	.83"	.43**	.67**	.66**	.26	.30"	.59"	.34"	.47**	.09	1					
Cu	.48"	.35"	.25	.22	.05	.04	.07	-20	$\cdot \mathbf{n}$.14	.19	.27	.65"	.16	.62**	.62**	.34"	.56**	.85"	.74**	.11	.27	.76**	1				
Cd	.43"	.26	.39"	.32"	.41"	.38"	.32*	.37"	.10	.15	.23	.26	-19	.01	.05	.063	.27	.19	-35"	.15	.55"	.41**	-13	.18	1			
Hg	.16	.56**	.47**	.52"	.25	.34"	.42**	.13	.02	20	.50"	.54"	.49"	.31"	.47**	.46**	.49"	.70**	.88"	.90"	.16	.48"	.55"	.89"	.15	1		
N	·44"	.50"	.39"	.68"	.75**	.59"	.69"	.64"	.53"	03	.63"	.54"	-26	-14	-13	-13	.20	.46**	.39"	.61**	.62**	.68**	10	.37"	.25	.57"	1	
Zn	.35"	.42**	.31*	.36"	.09	.15	.22	.08	.08	.20	.32"	.38"	.57"	.22	.53**	.52**	.41"	.62**	.87"	.84"	.03	.32*	.67**	.95"	.20	.95**	.48"	1

Hazard index (HI index). The health risks of surface water in the Karaca Pond were calculated using the US EPA risk evaluation models. With the help of Hazard Quotients (HQ), the intake of seven heavy metals to the body and its absorption through the skin were calculated (Tables 8 and 9). These values should not be more than 1. The hazard index (HI) values for children are higher than that of adults. However, the observed values for both children and adults were lower than 1. The total hazard index metal contamination is as follows: Pb >Cu >Cd >Ni> Hg >Fe.

Table 8. Hazard quotient (HQ) and hazard index (HI) for the trace levels in the Karaca Dam

	1									
	Adı	ult	Ch	uild	Ad	ult	Ch	nild	Adult	Child
Pb	1.67x10 ⁻³	7.40x10 ⁻⁶	1.87x10 ⁻³	1.65x10 ⁻⁵	1.19x10 ⁻³	1.75x10 ⁻⁵	1.33x10 ⁻³	3.92x10 ⁻⁵	1.12x10 ⁻³	1.37x10 ⁻³
Cd	3.29x10 ⁻⁴	3.43x10 ⁻⁵	3.68x10 ⁻⁴	7.59x10 ⁻⁵	6.57x10 ⁻⁴	$1.37 x 10^{-3}$	7.36x10 ⁻⁴	3.03x10 ⁻³	2.02x10 ⁻³	3.77x10 ⁻³
Fe	7.52x10 ⁻⁴	2.80x10 ⁻⁴	8.42x10 ⁻⁴	6.20x10 ⁻⁴	1,07x10 ⁻⁶	2.00x10 ⁻⁶	1.20x10 ⁻⁶	4.43x10 ⁻⁶	3.08x10 ⁻⁶	5.63x10 ⁻⁶
Cu	7.80x10 ⁻²	7.14x10 ⁻⁴	8.73x10 ⁻²	1.58x10 ⁻³	1.95x10 ⁻³	8.93x10 ⁻⁵	2.18x10 ⁻³	1.97x10 ⁻⁴	2.04x10 ⁻³	2.38x10 ⁻³
Zn	5.95x10 ⁻²	9.32x10 ⁻⁴	6.66x10 ⁻²	2.06x10 ⁻³	1.98x10 ⁻⁴	1.55x10 ⁻⁵	2.22x10-4	3.43x10 ⁻⁵	2.14x10-4	2.57x10-4
Ni	5.14x10 ⁻³	1.34x10 ⁻⁴	5.76x10 ⁻³	2.97x10 ⁻⁴	2.57x10 ⁻⁴	1.68x10 ⁻⁴	2.88x10 ⁻⁴	3.71x10 ⁻⁵	4.25x10 ⁻⁴	3.25x10 ⁻⁴

	Kpª	RFD ^a _{ingestion}	RFD^{a}_{dermal}	ABS (%) ^g
		µg/kg/day	µg/kg/day	%
Pb	$1x10^{-4}$	1.4	0.42	11.7
Cd	$1x10^{-3}$	0.5	0.025	5
Fe	$1x10^{-3}$	700	140	1,4
Cu	$1x10^{-3}$	40	8	57
Zn	6x10 ⁻⁴	300	60	20
Ni	2x10 ⁻⁴	20	0.8	4
Hg	$1x10^{-3}$	0.3	0.021	7
^a =	^b =			

Table 9. Reference done (RfD) in the Karaca dam.

The values obtained by Tesi et al., (2019) in the river delta showed similarity with our study. Pb, Cd, Fe, Cu, Zn, Ni, and Hg concentrations are quite low. Pb, which causes hematopoietic toxicity, was reported to be within limits. Kidney failure and immune deficiencies, along with bone injuries, are caused by Cd. Co is related to allergic dermatitis and rhinitis.

Conclusions

To determine the water quality of Karaca Dam Lake, which is used to provide drinking water, irrigation water to the sur-rounding area, samples were taken for 12 months from four stations representing the entire dam lake and waterquality analyses were carried out. To determine water quality, WQI was calculated by taking into account WHO (2011) standards. WQI values in the lake were found to be between 16.47 and 21.70. According to these results, Karaca Dam Lake surface water is in the 'Excellent class'. In the calculations made to determine irrigation water quality, the surface waters were found suitable in terms of SAR and sodium percent age content, whereas different properties were determined in terms of RSC throughout the year. According to SWQR, the water quality was found to be 'very good' in terms of DO, COD, BOD, nitrate and ammonium, but in terms of phosphate, it varied in range between the 'very good' and 'medium' classes . This suggests that there are no nutrient inputs from the surrounding domestic and agricultural areas that could damage the health of the eco-system. The limit values for heavy metals Cu, Zn and Fe were exceeded. The results of the present study showed that water quality parameters are within limits permitted by the WHO. Besides, this study is a pioneering effort for further investigation and risk assessment of trace metals present in water bodies located in regions with limited information in the Sinop province.

References

ABDULHUSSEIN F.M. (2018) Hydrochemical assessment of groundwater of dibdibba aquifer in al-zubair area, Basra, South of Iraq and its suitability for irrigation purposes. Iraqi Journal of Science, 59(1A):135–143. Doi: <u>10.24996/</u><u>ijs.2018.59.1A.15</u>

ALAM J.B., HOSSAIN A., KHAN S.K., BANIK B.K., ISLAM M.R., MUYEN Z., RAHMAN M.H. (2007) Deterioration of water quality of Surma river. Environmental Monitoring and Assessment, 134:233–242. Doi: <u>10.1007/</u> <u>\$10661-007-9612-7</u>

ALI E.M., KHAIRY H.M. (2016) Environmental assessment of drainage water impacts on water quality and eutrophication level of Lake Idku, Egypt. Environmental Pollution, 216:437–449. Doi: <u>10.1016/j.envpol.2016.05.064</u>

ALI J., KAZI T.G., TUZEN M., ULLAH N. (2017) Evaluation of mercury and physicochemical parameters in different depths of aquifer water of Thar coalfield, Pakistan. Environ. Sci. Pollut. Res.24(21):17731-17740. Doi: 10.1007/s11356-017-9291-z

ANWAR M.S.S., LATEEF S. AND SIDDIQI G.M. (2017) Bacteriological quality of drinking water in Lahore. Biomedica, 26(1):66–69.

APHA (1998) Standart methods for the ezamination of water and wastewater (20th ed.). Washington DC: American Public Health Association.

ARSHAD N., IMRAN S.(2017) Assessment of arsenic, fluoride, bacteria, and other contaminants in drinking water sources for rural communities of Kasur and other districts in Punjab, Pakistan. Environmental Science Pollution Reseseach, 24:2449–2463. Doi: 10.1007/s11356-016-7948-7

AWAIS M., ARSHAD M., SHAH S.H.H., ANWAR-UL-HAQ, M. (2017) Evaluating groundwater quality for irrigated agriculture: spatio-temporal investigations using GIS and geostatistics in Punjab, Pakistan. Arabian Journal Geoscience 10(23):510. Doi: 10.1007/s12517-017-3280-x

BANERJEE T., SRIVASTAVA R. (2009) Application of water quality index for assessment of surface water quality surrounding integrated industrial estate-Pantnagar. Water Science Techology 60(8):2041–2053. Doi: <u>10.2166/</u><u>wst.2009.537</u>

BELAL A.A.M., EL-SAWY M., DAR M.A. (2016) The effect of water quality on the distribution of macro-benthic fauna in Western Lagoon and Timsah Lake, Egypt. The Egyptian Journal of Aquatic Research, 42(4):437-448. Doi: <u>10.1016/j. ejar.2016.12.003</u>

BOYD C.E., TUCKER C.S. (1998) Pond aquaculture water quality management, Kluwer Academic Publishers, p. 700.

DAUD M.K., NAFEES M., ALI S., RIZWAN M., BAJWA R.A., SHAKOOR M.B., MALOOK I. (2017) Drinking water quality status and contamination in Pakistan. BioMed Research International. Doi: <u>10.1155/2017/7908183</u>

DEBELS P., FIGUEROA R., URRUTIA R., BARRA R., NIELL X. (2005) Evaluation of water quality in the Chilla'n river (Central Chile) using physicochemical parameters and a modified water quality index. Environmental Monitoring. Assessment, 110:301–322. Doi: 10.1007/s10661-005-8064-1

DONEEN L.D. (1975) Water quality for irrigated agriculture. In: Plants in Saline Environments, pp. 56–76. ISBN: 978-3-642-80929-3

EL-AZIZ S.H.A. (2017) Evaluation of groundwater quality for drinking and irrigation purposes in the north-western area of Libya (Aligeelat). Environmental Earth Sciences, 76 (4). Doi: <u>10.1007/s12665-017-6421-3</u>

FALOWO O.O., AKINDURENI Y., OJO O. (2017) Irrigation and drinking water quality index determination for groundwater quality evaluation in Akoko Northwest and Northeast Areas of Ondo State, Southwestern Nigeria. Am. Water Science and Engineering, 3(5):50. Doi: <u>10.11648/j.</u> <u>ajwse.20170305.11</u>

FATMI Z., AZAM I., AHMED F., KAZI A., GILL A.B., KADIR M.M. (2009) Core Group for Arsenic Mitigation in Pakistan. Health burden of skin lesions at low arsenic exposure through groundwaterinPakistan. Is river the source? Environmental Research, 109(5):575-581. Doi: <u>10.1016/j. envres.2009.04.002</u>

GHALIB H.B. (2017) Groundwater chemistry evaluation for drinking and irrigation utilities in east Wasit province, Central Iraq. Applied Water Science, 7:3447–3467. Doi: 10.1007/s13201-017-0575-8

GOLEKAR R.B., AKSHAY M., SHUBHAM J., SHUBHAM A., PATIL Y.M. (2017) Geochemical characteristics of water and its suitability for drinking and irrigation use in and around Warnanagar area of Kolhapur District (Maharashtra) India. Journal of Water Resources and Pollution Studies, 2(1):1-12.

GUO Q., WANG Y., MA T., MA R. (2007) Geochemical processes controlling the elevated fluoride concentrations in groundwaters of the Taiyuan Basin, Northern China. Journal of Geochemical Exploration, 93:1–12. Doi: <u>10.1016/j.</u> asej.2015.10.002

HUSSAIN Y., DILAWAR A., ULLAH S.F., AKHTER G., MARTINEZ-CARVAJAL H., HUSSAIN M.B., ASLAM A.Q, (2016a) Modelling the spatial distribution of arsenic in water and its correlation with public health, Central Indus Basin, Pakistan. Journal of Geoscience and Environment Protectiont, 4(2):18–25.

HUSSAIN Y., ULLAH S.F., HUSSAIN M.B., MARTINEZ-CARVAJAL H. AND ASLAM A.Q. (2016b) Protective capacity assessment of vadose zone material by geo-electrical method: a case study of Pakistan. International Journal of Geomechanics, 7(5):716–725.

HUSSAIN Y., ULLAH S.F., AKHTER G., ASLAM A.Q., AKHTER G., MARTINEZ-CARVAJAL H., IMNEISI I.B., AYDIN M. (2016c) Vulnerability assessment of an agrostressed aquifer using source-pathway-receptor model in GIS. Modeling Earth Systems and Environment, 3:595–604. Doi: 10.1007/s40808-017-0320-1

IMRAN M., JAHANZAIB S., ASHRAF A. (2017) Using geographical information systems to assess groundwater contamination from arsenic and related diseases based on survey data in Lahore, Pakistan. Arabian Journal of Geosciences, 10:450. Doi: 10.1007/s12517-017-3212-9

ISLAM A.T., SHEN S., BODRUD-DOZA M.D., RAHMAN M.S. (2017) Assessing irrigation water quality in Faridpur district of Bangladesh using several indices and statistical approaches. Arab. J. Geosci., 10(19):418. Doi: 10.1007/s12517-017-3199-2

JIANG L., YAO Z., LIU Z., WANG R., WU S. (2015) Hydrochemistry and its controlling factors of rivers in the source region of the Yangtze River on the Tibetan Plateau. Journal of Geochemical Exploration, 155:76–83. Doi: <u>10.1016/j.gexplo.2015.04.009</u>

KANGABAM RD, BHOOMINATHAN SD, KANAGARAJ S, GOVINDARAJU M (2017) Development of a water quality index (WQI) for the Loktak Lake in India. Applied Water Science, 7(6):2907–2918. Doi: <u>10.1007/s13201-017-0579-4</u>

KHANORANGA S.K. (2019) An assessment of groundwater quality for irrigation and drinking purposes around brick kilns in three districts of Balochistan province, Pakistan, through water quality index and multivariate statistical approaches. Journal of Geochemical Exploration, 197:14–26. Doi: <u>10.1016/j.gexplo.2018.11.007</u>

KIYMAZ S., KARADAVUT U. (2014) Application of multivariate statistical analysis in the assessment of surface water quality in Seyfe Lake, Turkey. Journal of Agricultural Sciences (Turkey), 20:152–163.

KUMAR P., BANSOD B.K.S., DEBNATH S.K., KUMAR P., GHANSHYAM C (2015) Index-based groundwater vulnerability mapping models using hydrogeological settings: a critical evaluation. Environmental Impact Assessment Review, 51:38–49. Doi: 10.1016/j.eiar. 2015.02.001

KUMAR P., THAKUR P.K., BANSOD B.K., DEBNATH S. (2018) Groundwater: a regional resource and a regional governance.Environ. Dev. Sustain. Environment Development and Sustainability, 20:1133–1151. Doi: 10.1007/s10668-017-9931-y

KUMAR P., THAKUR P.K. (2017a) Multi-criteria evaluation of hydro-geological and anthropogenic parameters for the groundwater vulnerability assessment. Environmental Monitoring Assessment, 189:564. Doi: <u>10.1007/s10661-017-6267-x</u>

KUMAR P., THAKUR P.K. (2017b) Groundwater: a regional resource and a regional governance.Environ. Dev. Sustain. Doi: <u>10.1007/s10668-017-9931-y</u>

KUMARI B., KUMAR V., SINHA A.K., AHSAN J., GHOSH A.K., WANG H., DEBOECK G. 2017 Toxicology of arsenic in fish and aquatic systems. Environ. Chem. Lett., 15(1):43–64. Doi: <u>10.1007/s10311-016-0588-9</u>

KUTLU B., KÜÇÜKGÜL A., DANABAŞ D. (2017) Annual and seasonal variation of nutrients and pigment content in Uzunçayir Dam Lake, Turkey (Eastern Anatolia). Elixir Environmental and Forestry, 112:48971-48974.

KUTLU B., TEPE R. (2019) Examination water quality of Karkamış Dam Lake. Turkish Journal of Agriculture Food Science and Technology, 7(3):458-466.

MORTUZAM.G., AL-MISNED F.A. (2017) Environmental contamination and assessment of heavy metals in water, sediments and shrimp of Red Sea Coast of Jizan, Saudi Arabia. J. Aquatic Pollution Toxicollogy 1(1).

MULLAI P., YOGESWARI M.K., OOMMEN B, SARAVANAKUMAR K., KATHIRESAN K. (2013) Monitoringof water quality parameters in UppanarRiver of Cuddalore District, Tamil Nadu State, India. Journal of Water Sustainability, 3(4): 179-192. MUMTAZ A., MIRJAT M.S., SOOMRO A. (2017) Assessment of drinking water quality status and its impact on health in Tandojam City. Journal of Basic and Applied Scientific research, 13:363–369. Doi: <u>10.6000/1927-</u> <u>5129.2017.13.60</u>

NASEEM S., HAMZA S., BASHIR E. (2010) Groundwater geochemistry of Winder agricultural farms, Balochistan, Pakistan and assessment for irrigation water quality. European Water, 31:21–32.

OBIEFUNA G.I., SHERIFF A. (2011) Assessment of Shallow Ground Water Quality of Pindiga Gombe Area Yola Area, NE, Nigeria for irrigation and domestic purposes. Research Journal of Environmental and Earth Sciences, 3:131-141. Retrieved from <u>http://maxwellsci.com/print/</u> rjees/v3-131-141.pdf

ONGLEY E.D., BIRKHOLZ D.A., CAREY J.H., SAMOILOFFM.R.(1988)Iswaterarelevantsamplingmedium for toxic chemicals? An alternative environmental sensing strategy. Journal of Environmental Quality,17(3):391407. Doi: 10.2134/jeq1988.00472425001700030008x

PADHI S., RANGARAJAN R., RAJESHWAR K., SONKAMBLE S., VENKATESAM V. (2017) Assessment of Hydro-geochemical evolution mechanism and suitability of groundwater for domestic and irrigation use in and around Ludhiana city, Punjab, India. Journal of Indian Geophysical Union, 4:260–270. Doi: <u>10.1080/10807039.2017.1292117</u>

PATEL K.S., SAHU B.L., DAHARIYA N.S., BHATIA A., PATEL R.K., MATINI L., BHATTACHARYA P. (2017) Groundwater arsenic and fluoride in Rajnandgaon District, Chhattisgarh,northeastern India. Applied Water Science, 7:1817-1826. Doi: <u>10.1007/s13201-015-0355-2</u>

PHUNG D., HUANG C., RUTHERFORD S., DWIRAHMADI E., CHU C., WANG X., NGUYEN M., NGUYEN N.H., DO C.M., NGUYEN T.H., DINH T.A.D. (2015) Temporal and spatial assessment of riversurfacewater quality using multivariate statisticaltechniques: a study in Can Tho City, a MekongDelta area, Vietnam. Environmental Monitoring and Assessment, 187(5):229. Doi: <u>10.1007/</u> <u>\$10661-015-4474-x</u>

PODGORSKI J.E., EQANI S., HEQING S., BERG M. (2016) Geospatial modeling of widespread arsenic contamination in unconfined, high-pH aquifers in Pakistan. In:AGU Fall Meeting Abstracts.B14A-08.

RAFIQUE T., NASEEM S., BHANGER M.I. AND USMANI T.H. (2008) Fluoride ion contamination in the groundwater of Mithi sub-district, the Thar Desert, Pakistan. Environmental Geology, 56:317–326. Doi: <u>10.1007/</u> <u>s00254-007-1167-y</u>

RAJU N.J. (2007) Hydrogeochemical parameters for assessment of groundwater quality in in the upper Gunjanaeru River basin, Cuddapah District, Andhra Pradesh, South India. Environmental Geology, 52:1067–1074. Doi: 10.1007/s00254-006-0546-0

RAJU N.J., SHUKLA U.K., RAM P. (2011) Hydrogeochemistry for the assessment of groundwater quality in Varanasi: a fast-urbanizing center in Uttar Pradesh, India. Environmental Monitoring and Assessment, 173:279– 300. Doi: 10.1007/s10661-010-1387-6

RAPANT S., CVEČKOVA V., FAJČIKOVA K., SEDLAKOVA D., STEHLIKOVA B. (2017) Impact of Calcium and Magnesium in Groundwater and Drinking Water on the Health of Inhabitants of the Slovak Republic. International Journal of Environmental Research and Public, 14 (3):278. Doi: <u>10.3390/ijerph14030278</u>

RASOOL A., FAROOQI A., MASOOD S., HUSSAIN K. (2016) Arsenic in groundwater and its health risk assessment in drinking water of Mailsi, Punjab, Pakistan. Human and Ecological Risk Assessment, 22(1):187–202. Doi: 10.1080/10807039.2015.1056295

Rasool A., Xiao T., Farooqi A., Shafeeque M., Liu Y., Kamran M.A. and Eqani S. (2017) Quality of tube well water intended for irrigation and human consumption with special emphasis on arsenic contamination at the area of Punjab, Pakistan. Environmental Geochemistry and Health, 9(4):847–863. Doi: 10.1007/s10653-016-9855-8

RUPAKHETI D., TRIPATHEE L., KANG S., SHARMA C.M., PAUDYAL R., SILLANPÄÄ M. (2017) Assessment of water quality andhealth risks for toxic trace elements in urban Phewa andremote Gosainkunda lakes, Nepal., Human and Ecological Risk Assessment: An International Journal, 23(5):959–973. Doi: 10.1080/10807039.2017.1292117

SALLAM G.A.H., ELSAYED E.A. (2018) Estimating relations between temperature, relative humidity as independed variables and selected water quality parameters in Lake Manzala, Egypt. Ain Shams Engineering Journal, 9(1):1–14. Doi: 10.1016/j.asej.2015.10.002

SENER S., SENER E., DAVRAZ A. (2017) Evaluation of water quality using water quality index (WQI) method and GIS in Aksu River (SW-Turkey). Science of the Total Environment, 584–585:131–144. Doi: <u>10.1016/j.</u> scitotenv.2017.01.102 SING K.S.W., EVERETT D.H., HAUL R.A.W., MOSCOU L., PIEROTTI R.A., ROUQUEROL J., SIEMIENIEWSKA T. (2008) Reporting Physisorption Data for Gas/Solid Systems. In:Handbook of Heterogeneous Catalysis. Wiley-VCH Verlag GmbH and Co. KGaA, pp.1217–1230. Doi: 10.1002/9783527610044.hetcat006

SINGARAJA C., CHIDAMBARAM S., ANANDHAN P., PRASANNA M.V., THIVYA C., THILAGAVATHI R., SARATHIDASAN J. (2014) Hydrochemistry of groundwater in a coastal region and itsrepercussion on quality, a case study - Thoothukudi district, Tamil Nadu, India. Arabian Journal of Geosciences, 7(3):939–950. Doi: <u>10.1007/s12517-012-0794-0</u>

SOOMRO F., RAFIQUE T., MICHALSKI G., ALI S.A., NASEEM S., KHAN M.U. (2017) Occurrence and delineation of high nitrate contamination in the groundwater of Mithi sub-district, Thar Desert, Pakistan. Environmental Earth Science, 76 10):355. Doi: <u>10.1007/s12665-017-6663-0</u>

SRINIVAS Y., AGHIL T.B., OLIVER D.H., NAIR C.N., CHANDRASEKAR N. (2017) Hydrochemical characteristics and quality assessment of groundwater along the Manavalakurichi coast, Tamil Nadu, India. Applied Water Science, 7(3):1429–1438. Doi: <u>10.1007/s13201-015-0325-8</u>

SUBRAMANI T., RAJMOHAN N., ELANGO L. (2009) Groundwater geochemistry and identifica-tion of hydrogeochemical processes in a hard rock region, Southern India. Environmental Monitoring Assessment, 162(1–4):123–137. Doi: 10.1007/s10661-009-0781-4

SWQR (2016) Turkey's Ministry of Forestry and Water Affairs Surface Water Quality Regulations. <u>http://www. resmigazete.gov.tr/eskiler/2016/08/20160810-9.htm.%20</u> <u>Accessed%2021March%202018</u>

TEPE R. KUTLU B. (2019) Examination water quality of Karkamış Dam Lake. Turkish Journal of Agriculture Food Science and Technology, 7(3):458-466. Doi: <u>10.24925/</u> turjaf.v7i3.458-466.2409

TESI G.O., TESI J.A., OGBUTA A.A., INIAGHE P.O., ENETE C.I. (2019) Assessment of effect of sandmining activities on physicochemical properties and metal concentrations of surface water of warri river, niger delta, Nigeria. Fudma Journal of Science, 3(1):72-83. ISSN: 2645-2944

VAROL S., DAVRAZ A. (2015) Evaluation of the groundwater quality with WQI (Water Quality Index)

and multivariate analysis: a case study of the Tefenni plain (Burdur/Turkey). Environmental Earth Science, 73:1725–1744. Doi: <u>10.1007/s12665-014-3531-z</u>

VASANTHAVIGAR M., SRINIVASAMOORTHY K., VIJAYARAGAVAN K., GANTHI R.R., CHIDAMBARAM S., ANANDHAN P., VASUDEVAN S.(2010) Application of water quality index for groundwater quality assessment: Thirumanimuttar sub-basin, Tamilnadu, India. Environmental Monitoring and Assessment, 17181(4):595-609. Doi: <u>10.1007/s10661-009-1302-1</u>

VEGA M., PARDO R., BARRATO E., DEBAN L. (1998) Assessment of seasonal and polluting effects on the quality of river water by exploratory data analysis. Water Research, 32:3581–3592. Doi: <u>10.1016/S0043-1354(98)00138-9</u>

WHO (2011) WHO Guidelines for Drinking-water Quality, fourth ed. World Health Organization.

WU Z., WANG X., CHEN Y., CAI Y., DENG J. (2018) Assessing river water quality using water quality index in Lake Taihu Basin. China. Science of the Total Environment, 612:914-922. Doi: <u>10.1016/j.scitotenv.2017.08.293</u> YISA J., JIMOH T. (2010) Analytical studies on Water Quality Index of river Landzu. American Journal of Applied Sciences, 7(4):453-458. Doi: 10.3844/ajassp.2010.453.458

YU C., LI Z., YANG Z. (2015) A universal calibrated model for the evaluation of surface water and groundwater quality : Model development and a case study in China 163,20-27. Journal of Environmental Managemen, 163(1):20-27. Doi: 10.1016/j.jenvman.2015.07.011

ZAFAR S., KHAN A., ULLAH H., KHAN M.S., KHAN I., HAMEED A., YASMEEN G. (2017) Assessing impact of effluent discharge on irrigation water quality in southern region. Environmental Monitoring and Assessment, 189(4):156. Doi: <u>10.1007/s10661-017-5868-8</u>

ZHANG Z., WANG J.J., ALI A., DELAUNE R.D. (2016) Heavy metal distribution and waterquality characterization ofwater bodies in Louisiana's Lake Pontchartrain Basin, USA. Environmental Monitoring and Assessment, 188(11):628. Doi: 10.1007/s10661-016-5639-y